

## **A graphical non-parametric hydrologic alteration test using flow duration curves**

Mohamed H. Mowafy, Department of Environmental Resources Engineering, SUNY College of Environmental Science and Forestry, Syracuse, NY 13210.

\*Charles N. Kroll, Department of Environmental Resources Engineering, SUNY College of Environmental Science and Forestry, Syracuse, NY 13210, [cnkroll@esf.edu](mailto:cnkroll@esf.edu).

Richard M. Vogel, Department of Civil and Environmental Engineering, Tufts University, Medford, MA 02155.

\*Corresponding author

1 **A graphical non-parametric hydrologic alteration test using flow duration curves**

2 **Abstract**

3 Hydrologic alteration results from a variety of natural and human forces including climate  
4 change and variability, water use and water infrastructure, and land cover alterations. The  
5 assessment of statistically significant hydrologic alteration is a critical challenge. One common  
6 tool to aid in this assessment is an annual flow duration curve (FDC), a plot of daily streamflow  
7 versus its probability of exceedance in a given year. Utilizing annual FDCs, the Modified  
8 Mood’s Median Test (MMMT) was applied to percentiles of the median annual FDCs at 122  
9 disturbed sites in the conterminous U.S. A simple graphical tool based on the MMT was  
10 developed to illustrate the significance of hydrologic alteration over a wide range of streamflow  
11 conditions at each site. Hydrologic disturbance was mostly caused by dams, though other forms  
12 of alteration were explored. In addition, 308 minimally regulated HCDN sites were investigated  
13 to assess regional climatic drivers of hydrologic alteration. Results indicate that dams can have  
14 varying effects on the hydrologic regime, but generally impact low flows more often than high  
15 flows in the U.S. Urbanization and water transfers can also have large impacts on hydrologic  
16 regimes. At HDCN sites, we observe significant changes to hydrologic regimes in many regions,  
17 especially to low and median flows in the upper midwestern and eastern U.S. A graphical tool  
18 based on the MMT was developed to show the significance of hydrologic alteration over a  
19 range of streamflow conditions; this tool is publicly available for other applications.

20

21 Key Words: Hydrologic alteration, flow duration curve, streamflow data, non-parametric  
22 hypothesis tests, median.

23 **1.0 Introduction**

24 Hydrologic systems evolve in response to a variety of natural and human forces such as climate  
25 change and variability and change, water use and water infrastructure, and land cover alterations.  
26 Such changes in hydrologic systems impact socioeconomic, ecological, and climate systems at  
27 multiple scales, leading to a coevolution of these interlinked systems (Vogel et al., 2015).  
28 Hydrologic disturbance can result from either intended or unintended activities. Rivers and  
29 streams system can be disturbed through flow regulation to either mitigate the outcome of  
30 extreme events (e.g., floods, droughts) or to better make use of these resources (e.g., water  
31 supply, irrigation, hydropower) (Malmqvist and Rundle, 2002; Richter et al., 1996). While river  
32 flow regulation often results from the construction and operation of dams (Nilsson et al., 2005;  
33 Poff et al., 2007), including off-stream reservoirs (Hecht et al., 2014), streams can also be  
34 disturbed due to climate and land use changes, interbasin water transfers, and inputs to streams  
35 such as via wastewater treatment plant discharges (Bradford et al., 2018). The resulting  
36 hydrologic alteration can cause loss of floodplains and wetlands, habitat fragmentation and  
37 ecological degradation, reduction in water quality, and the loss of ecosystem services (Bunn and  
38 Arthington, 2002; Magilligan and Nislow, 2005; Postel et al., 1996; Rosenberg et al., 2000;  
39 Sabater et al., 2018; Yusa et al., 2015). Although predicting the impact of hydrologic alteration  
40 on ecosystem services may be challenging, a recent review article by Jumani et al. (2020)  
41 concludes that improved assessment of the significance of hydrologic alteration is needed to  
42 better inform both water and ecological management decisions.

43

44 To assess hydrologic alterations, various metrics and indices have been developed (Dudley et al.,  
45 2020; Gao et al., 2009; Poff et al., 2010; Richter et al., 1996; Yang et al., 2017, 2008). Richter et

46 al. (1996) proposed the Indicators of Hydrologic Alteration (IHA), which consists of 33  
47 streamflow statistics that characterize a wide range of hydrologic conditions. Poff et al. (2010)  
48 developed the Ecological Limits of Hydrologic Alteration (ELOHA) framework providing  
49 empirical relationships between flow alteration and ecological responses. When using such  
50 metrics, assessing the significance of alteration can be challenging. Individual statistics which  
51 comprise these metrics often capture only one portion of the hydrologic cycle or only one aspect  
52 of the hydrologic phenomenon which impacts the system (e.g., flows in a specific month). For  
53 instance, Dudley et al. (2020) utilized the annual 7-day low streamflow to detect trends within  
54 human-impacted basins in the U.S. Unfortunately, many of the advocated metrics and indices are  
55 plagued by both redundancy and correlation (Olden and Poff, 2003), which has led to the  
56 development of reduced sets of indices. Gao et al. (2009) used principal component analysis to  
57 produce a smaller set of independent and characteristic hydrologic indicators which could more  
58 fully represent a wider range of streamflow conditions impacted by hydrologic alteration. Yang  
59 et al. (2008) used a data mining approach to identify the most ecologically relevant hydrologic  
60 indicators (ERHIs) from the IHA. Yang et al. (2017) developed a Criteria of Importance Through  
61 Intercriteria Correlation (CRITIC) algorithm to produce more representative indicators of  
62 hydrologic alteration. Vogel et al. (2007) introduced integrated indicators of hydrologic  
63 alteration referred to as seasonal ecodeficit/ecosurplus for evaluating the impact of reservoir  
64 regulation on ecological flow regimes. While these proposed approaches address some of the  
65 issues of redundancy and correlation among metrics, it is still unclear how many or which  
66 indices are needed to fully reflect hydrologic alteration. There is a need for a convenient tool that  
67 could help to evaluate the significance of hydrologic alteration resulting from a wide range of  
68 types and causes of streamflow alteration.

69

70 One common tool used to assess hydrologic alteration is a flow duration curve (FDC). A FDC is  
71 a plot of streamflow versus its probability of exceedance (Vogel and Fennessey, 1994). The FDC  
72 captures the entire range of streamflow magnitudes, and thus integrates the complete streamflow  
73 regime into a single tool. FDCs have been utilized for a wide range of applications, including but  
74 not limited to hydropower design, habitat assessment, flood abatement, water quality evaluation,  
75 wasteload allocations, inundation mapping and for comparative hydrologic assessments  
76 (Castellarin et al., 2013; Vogel and Fennessey, 1995). In one of the most widely cited articles on  
77 the application of FDC's, Brown et al. (2005) document the value of FDC's in paired catchment  
78 studies for "gaining a greater understanding of the impact of vegetation on the distribution of  
79 daily flows." In a recent review of applications of FDCs over the period 2000-2020, Leong and  
80 Yokoo (2021) highlighted an "enormous volume of literature available on FDC studies within  
81 this time frame" and they highlighted the need for future research to improve our ability to  
82 incorporate anthropogenic influences on the behavior of FDCs. Many hydrologists, scientists and  
83 water managers would not make water resource decisions without reference to a FDC. Even  
84 statisticians consider graphical displays as the single most effective and robust statistical tool  
85 (Wainer, 1990). Importantly, FDCs are an effective graphical instrument for conveying  
86 information regarding a wide range of water resource problems; hence, they provide an effective  
87 medium for communication between water resource engineers, lawyers, managers, planners,  
88 politicians, and others. It is for these reasons that our goal is to develop a graphical hypothesis  
89 test based on a FDC. A FDC is also a convenient tool for the detection and attribution of  
90 hydrologic change. For instance, if a dam and associated reservoir is introduced, one might  
91 expect a flattening of the FDC downstream of the dam; water withdrawals or reductions in

92 precipitation would generally result in a lowering of the entire FDC (Atieh et al., 2017; Moyle  
93 and Mount, 2007; Vogel et al., 2007). FDCs are part of a river's *streamflow signature* that can  
94 represent both the natural state of rivers and their responses to hydrologic alteration.

95

96 It is often challenging to determine whether alterations to the hydrologic system are significant  
97 or are instead simply the consequences of the natural variability of streamflow under stationary  
98 conditions (Allan, 2004). Kroll et al. (2015) used shifts in annual FDCs to develop four  
99 hypothesis tests of hydrologic alteration based on different methods of bootstrap resampling.

100 Using dam sites in the U.S. which were assumed to have undergone hydrologic alteration, they  
101 found a confidence interval test (which is more accurately termed a "tolerance interval test") of  
102 the average annual FDC to have the highest power, among the four tests considered.

103 Unfortunately, that test only examined the most extreme shifts in the mean annual FDC and did  
104 not consider hydrologic alteration across a range of streamflow conditions, a goal of this study.

105 In addition, the test proposed by Kroll et al. (2015) was generally based on the condition when  
106 the post-alteration streamflow record was much shorter than the pre-alteration streamflow record,  
107 allowing for bootstrap resampling of the pre-alteration streamflow to develop the critical value of

108 the hypothesis test. Hecht et al. (2020) assessed hydropower-ecosystem tradeoffs by examining  
109 the decrease of the 5<sup>th</sup> percentile and the increase of the 95<sup>th</sup> percentile of annual FDCs,

110 determining the type-1 and type-2 errors of a Mann-Whitney-Wilcoxon test (MWW). While the

111 hypothesis tests introduced by Kroll et al. (2015) and Hecht et al. (2020) do provide an

112 evaluation of the significance of hydrologic alteration, the resulting tests are generally

113 complicated to understand, implement and explain to water managers. A more general

114 hypothesis test of hydrologic alteration that is easy to interpret and communicate and could be

115 applied to a wider set of streamflow conditions is needed. We exploit the widely understood and  
116 attractive graphical features of FDCs to enable the development of such a hypothesis test as well  
117 its ease of implementation and acceptance in practice. One approach to the development of  
118 hypothesis tests for annual FDC's would be to exploit confidence (or more correctly, tolerance)  
119 intervals for annual FDC's introduced by Vogel and Fennessey (1994), reviewed by Castellarin  
120 et al. (2007) and extended more generally by Serinaldi (2011). Serinaldi (2011) introduced an  
121 attractive and general analytical approach to develop confidence intervals for annual flow  
122 duration curves, which could be adapted to construct hypothesis tests of hydrologic alteration.  
123 However, since his approach is based on index FDC's and requires assumptions concerning the  
124 distribution of streamflow, we did not adapt this approach here for our purposes.

125

126 Due to inherent problems in parametric hypothesis tests when distributional assumptions are  
127 violated (Ernst and Albers, 2017; Williams et al., 2013), nonparametric approaches are often  
128 preferred (Helsel et al., 2019), especially when dealing with daily streamflow series which  
129 exhibit extremely complex tail behavior (Blum et al., 2017). A non-parametric modification of  
130 Mood's median hypothesis test proposed by Fligner and Rust (1982) is explored here. The  
131 median (as opposed to the mean) better captures the central tendency of skewed data (Ames,  
132 2006; Helsel et al., 2019), and FDCs based on daily streamflows are known to exhibit  
133 extraordinary levels of skewness (Ye et al., 2021). Vogel and Fennessey (1994) introduced the  
134 median annual FDC which represents the FDC in a "typical" or median year. Our hypothesis  
135 testing approach evaluates whether or not flow variations within a typical or median year differ  
136 before and after a potential hydrologic disturbance. A case study is performed using 122 gaged  
137 U.S. streamflow records that are impacted by different types of disturbance. Although most of

138 the sites included in this study are streams with dams, we also explore other types of disturbance  
139 including urbanization, wastewater treatment plant (WWTP) effluent discharge, and interbasin  
140 water transfer. We also explore our hypothesis test using 308 minimally disturbed U.S. gaged  
141 sites to assess impacts of potential regional climatic changes. Our hypothesis test is also  
142 illustrated using a simple graphical tool to enable visualization of hydrologic alteration across the  
143 entire range of streamflow conditions. This graphical tool, which was created in R, produces a  
144 graphical representation of hydrologic alteration that is easy to implement, understand and  
145 communicate, should aid water planners and managers when addressing potential hydrologic  
146 alterations, and is posted on Github (<https://github.com/cnkroll/MMMT.git>) for others to use.  
147 Our overall goal is to introduce a nonparametric graphical hypothesis test which should be useful  
148 in most applications of annual FDC's, a considerable challenge given the now widespread  
149 application of annual FDC's in hydrology.

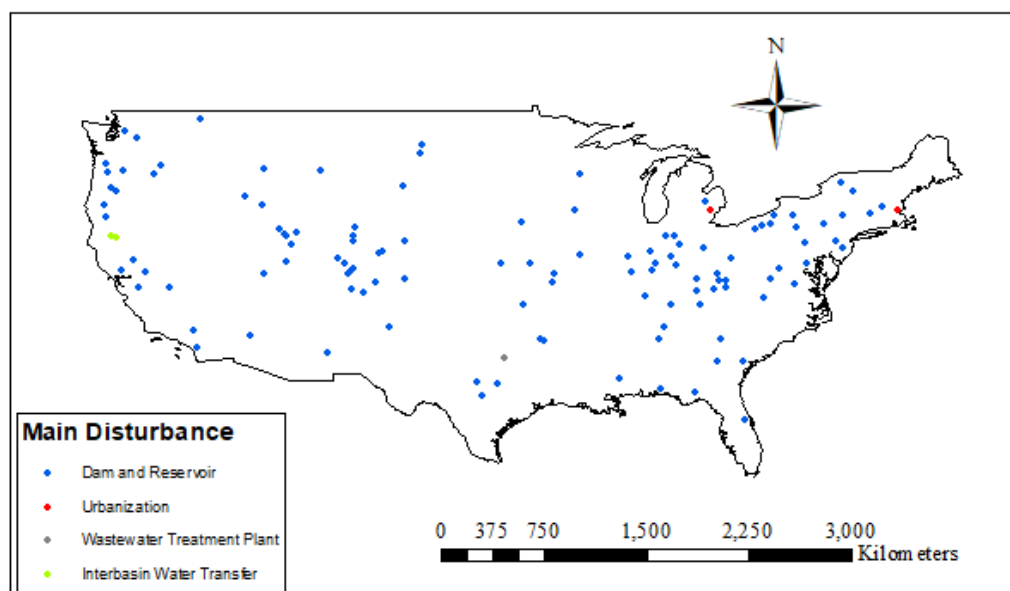
## 150 **2.0 Study Area and Methodology**

### 151 **2.1. Study area**

152 Case studies at USGS streamflow gages are implemented to evaluate the proposed hypothesis  
153 test. We employ 3 sets of gages: (1) 117 USGS streamflow gages with dams previously used by  
154 Kroll et al. (2015), (2) 5 USGS streamflow gages which represent other forms of potential  
155 hydrologic disturbance, and (3) 308 minimally regulated sites from the USGS Hydro-Climatic  
156 Data Network (HCDN). All sites have at least 15 water years of daily observations both before  
157 and after the “disturbance” year. Note that a disturbance could be an anthropogenic intervention  
158 such as a dam, urbanization, or interbasin water transfer, or it could be a potential climatic  
159 change point. Figure 1 shows the locations of the 122 potentially disturbed sites, while Figure 6  
160 shows the location of the 308 minimally regulated HCDN sites considered. Mowafy (2021,



161 Appendix 1) summarizes the 122 disturbed sites, providing their drainage area, hydrologic  
162 disturbance index (HDI), type and year of disturbance and the start and end year of the  
163 streamflow record. The HDI is a composite index of hydrologic disturbance which ranges from 0  
164 (no disturbance) to approximately 40 (heavily disturbed) (Falcone et al., 2010). The HDI is  
165 composed of seven disturbance metrics: major dam density, change in dam storage, percent  
166 canals and artificial paths, distance to National Pollutant Discharge Elimination Sites, freshwater  
167 withdrawals, road density, and landscape fragmentation.



168  
169 **Figure 1:** Locations of the 122 USGS stream gages with potential disturbance employed in this  
170 analysis (minimally regulated HCDN sites' locations depicted in Figure 6).

171

## 172 **2.2. Modified Mood's Median Test Methodology (MMMT)**

173 At each site a continuous daily streamflow record was separated into water years (October 1<sup>st</sup> to  
174 September 30<sup>th</sup>) and divided into pre- and post-alteration samples. For the sites with dams, the 5  
175 years before and after the disturbance were omitted from the record to minimize the impact of  
176 dam construction and reservoir filling. The median annual FDC was constructed for 199  
177 percentiles ( $p = 0.005, 0.01, 0.015, \dots, 0.995$ ) employing the nonparametric quantile estimator

178 introduced by Harrell and Davis (1982) in which each percentile of the FDC is computed from a  
179 linear combination of all the ordered observations. Vogel and Fennessey (1994) evaluated four  
180 different estimators for FDC quantiles and found that among those estimators, the estimator  
181 proposed by Harrell and Davis (1982) and given in equation 5a of Vogel and Fennessey (1994)  
182 provided smoother estimators for small samples than alternative estimators based on only a few  
183 of the ordered observations. Below we describe a hypothesis test of medians for each percentile  
184 of pre- and post-disturbance annual FDCs.

185

186 Let  $X_1, \dots, X_m$  and  $Y_1, \dots, Y_n$  denote independent random samples (from a pre-disturbance and  
187 post-disturbance annual FDC percentile), with sample sizes  $m$  and  $n$ , respectively. In addition, let  
188  $Z_1, \dots, Z_N$  denote the combined sample where  $N = m+n$ . The true (population) medians for both  
189 populations are assumed to be  $M_1$  and  $M_2$ , so that the cumulative density functions (CDFs) are  
190  $F_1(M_1) = F_2(M_2) = 0.5$ . Of concern is testing:

191 
$$H_0: M_1 - M_2 = 0$$

192 
$$\text{versus } H_a: M_1 - M_2 \neq 0$$

193 Friedman (1937) proposed an analysis of ranked data instead of an analysis of variance to  
194 perform the above hypothesis test to avoid the assumption of normality. Friedman's ranked-  
195 based test statistic approaches a Chi-square distribution as the sample size increases (Friedman,  
196 1937). Based on this approach, the Mood's median test was developed by ranking the combined  
197 set of observations and utilizing the sum of squared deviations of rank totals from their  
198 corresponding expected values as a test statistic (Brown and Mood, 1951). Mood's test statistic,  
199  $T$ , is the number of observations from one of the samples (here assumed to be  $Y_i$ ) that are less  
200 than or equal to the combined sample median,  $\hat{M}$ . An attractive feature of this test is that under

201 the null hypothesis, when the medians of both distributions are equal, the test statistic only  
 202 depends on the ranks of the observations and is thus a distribution free or nonparametric test.

203  
 204 In the derivation of their test, Brown and Mood (1951) assumed that both populations have the  
 205 same distributional shape, a poor assumption for this study. Instead, we utilize a modification to  
 206 the Mood's test that was proposed by Fligner and Rust (1982) to avoid the assumption of  
 207 similarly shaped distributions. Under the null hypothesis (equal medians),  $(n^*(T-1))^{1/2}$  will have  
 208 a limiting normal distribution with a mean of 0 and a variance of  $\sigma^2$ :

$$209 \quad \sigma^2 = \frac{1}{4} \rho (1 + \rho R^2) / (1 + \rho R)^2 \quad (1)$$

210 where  $\rho = \lim(m/n)$  as  $m \rightarrow \infty$  and  $n \rightarrow \infty$  and  $R = f_1(M)/f_2(M)$ , where  $f_1(M)$  and  $f_2(M)$  are the  
 211 probability density functions of the medians of the pre- and post-disturbance populations,  
 212 respectively. Under the Mood's original median test assumption that both records exhibit the  
 213 same distributional shape,  $R$  would equal 1 and the variance in (1) simplifies to:

$$214 \quad \sigma^2 = \frac{1}{4} \rho / (1 + \rho) \quad (2)$$

215 When the samples do not have the same distributional shape, a new estimator of  $\sigma^2$  based on a  
 216 range of order statistics centered on the sample median was proposed by Fligner and Rust (1982).

217 The lower and upper bounds of this range are  $L = Z_{N-b_N}$  and  $U = Z_{b_N}$ , respectively, where  $b_N$   
 218 assigns the rank location of these bounds using:

$$219 \quad b_N = \left[ \frac{1}{2}(N+1) + N^{\frac{1}{2}} \right] \quad (3)$$

220 where the square brackets represent the greatest integer function and  $Z_i$  denotes the  $i^{\text{th}}$  ordered  
 221 value of  $Z$  from the combined sample of length  $N=m+n$ .  $R$  is then estimated by:

$$222 \quad \hat{R} = (F_1(U) - F_1(L)) / (F_2(U) - F_2(L)) \quad (4)$$

223 where  $F_1$  and  $F_2$  are estimated by linearly interpolating across the pre- and post-disturbance cdfs  
 224 using a Weibull plotting position. If either L or U falls out of the pre- or post-disturbance sample  
 225 range for a given percentile, it is assigned to be either the smallest or the largest streamflow  
 226 value of the corresponding sample, and that value is also employed to calculate  $F_i(U)$  or  $F_i(L)$  for  
 227 both samples. Fligner and Rust (1982) proposed the MMMT test statistic:

$$228 \quad \hat{T} = n^{\frac{1}{2}} (T - \frac{1}{2}) / \hat{\sigma} \quad (5)$$

229 where:

$$230 \quad \hat{\sigma}^2 = \frac{1}{4} p (1+p \hat{R}^2) / (1+p \hat{R})^2 \quad \text{if } F_1(U) - F_1(L) > 0 \text{ and } F_2(U) - F_2(L) > 0 \quad (6)$$

$$231 \quad \hat{\sigma}^2 = \frac{1}{4} p \quad \text{if either } F_1(U) - F_1(L) = 0 \text{ or } F_2(U) - F_2(L) = 0$$

232 Where  $p = m/n$ . Note that for the case in equation 6 where either  $F_1(U) - F_1(L) = 0$  or  $F_2(U) -$   
 233  $F_2(L) = 0$  we have modified an apparent typographic error ( $\hat{\sigma}^2 = \frac{1}{4}$ ) in Fligner and Rust (1982 p.  
 234 223). The null hypothesis is rejected when  $|\hat{T}| \geq t_{N, \frac{\alpha}{2}}$  where  $t_{N, \frac{\alpha}{2}}$  is the  $(1 - \frac{\alpha}{2}) * 100\%$  percentile of  
 235 a Student T distribution with  $N=(m+n)$  degrees of freedom. The p-value for a two-tailed test can  
 236 be determined by:

$$237 \quad \text{p-value} = 2P(T_N \geq |\hat{T}|) \quad (7)$$

238 In some situations when nearly all of the values of  $F_1$  are greater than  $F_2$  (or vice versa), L may  
 239 become greater than U when either L or U falls out of the pre- or post-disturbance sample range  
 240 for a given percentile (from equation 6). For this case,  $\hat{T}$  is set to -3 and p-value is set to 0.005 as  
 241 defaults.

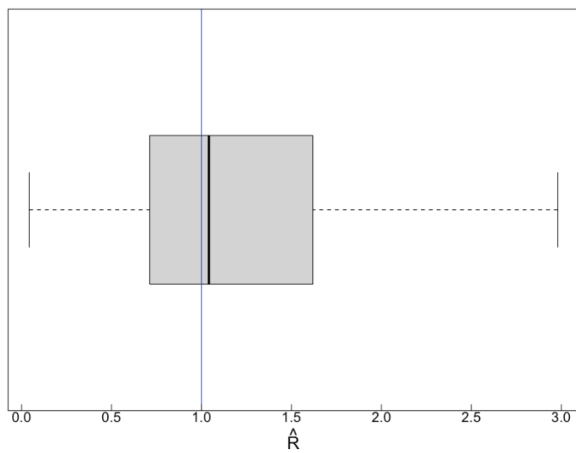
### 242 **3.0 Graphical Hypothesis Test Results**

243 In Section 3.1, the MMMT is performed at the 117 sites with dams. In Section 3.2, a graphical  
 244 tool is developed to visually assess hydrologic alteration at percentiles of median annual FDCs,

245 and selected dam sites from our dataset are examined to better understanding the impacts of dam  
246 and reservoir systems on resulting hydrologic disturbance. In Section 3.3, sites with other forms  
247 of disturbance are examined. Finally, in Section 3.4, the test is performed at HCDN sites using  
248 1970 as a climatic “break-point”.

### 249 3.1. Performance of MMMT at Selected U.S. Dam Sites

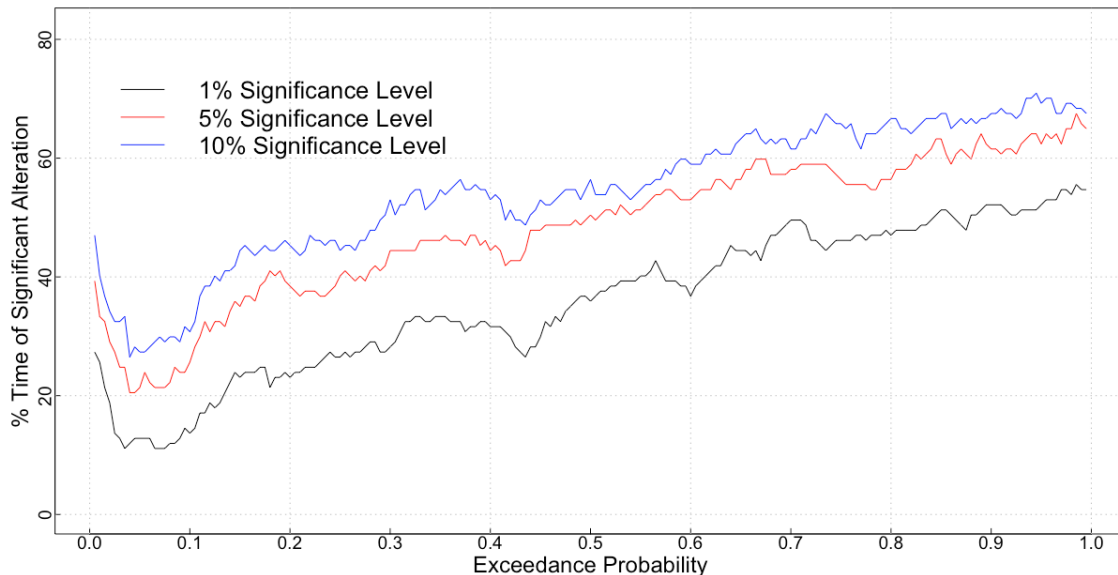
250 If  $\hat{R}$  from equation (4) is equal to 1, then MMMT defaults to the Mood’s Median Test. Using the  
251 117 sites with dams,  $\hat{R}$  was calculated for each percentile, and Figure 2 presents a box plot of  $\hat{R}$   
252 across all sites at all percentiles. Across these sites and percentiles, there is a wide range of  $\hat{R}$ , and  
253 a t-test assessing  $H_0: R = 1$  versus  $H_a: R \neq 1$  indicates that  $R$  is significantly different than 1 across  
254 all sites (p-value  $\ll 0.01$ ).



255  
256 **Figure 2:** Box plot of  $\hat{R}$  values across 117 U. S. dam sites at all percentiles.  
257

258 To evaluate the outcome of applying the MMMT across all the dam sites, Figure 3 shows the  
259 percentage of dam sites with a significant alteration for a given percentile at three different  
260 hypothesis test significance levels (1%, 5% and 10%) based on the pre- and post-alteration annual  
261 FDCs. In general, there was a general increase in the percentage of sites with significant changes  
262 in the median annual FDC percentile as the exceedance probability increases. At significance  
263 levels of 10%/5%/1%, approximately 20%/25%/35% of the sites experienced a significant

264 alteration at extreme high flows (0.5<sup>th</sup> to 5<sup>th</sup> percentiles) and approximately 55%/65%/75% of the  
 265 sites had significant alteration at extreme low flows (95<sup>th</sup> to 99<sup>th</sup> percentiles), respectively,  
 266 indicating that dams more often significantly alter low flows than high flows in the U.S. This  
 267 makes sense given that streamflow alterations would generally have a larger impact on the lowest  
 268 streamflows. In addition, at the lowest exceedance probabilities (highest flows), we see a slight  
 269 increase in the percentage of sites with significant alteration. Dams are often installed for flood  
 270 control, reducing the impact of the largest streamflows. Note that a comparison (not shown here)  
 271 was made between the MMMT and a similar test based on the ratio of medians using the method  
 272 developed by Bonett and Price (2020). Results indicated that the 2 methods had a similar  
 273 percentage of alteration across the dam sites shown in Figure 1 (Mowafy, 2021). Since the Bonett  
 274 and Price method requires a logarithmic transformation, it cannot easily handle flows recorded as  
 275 0 (unlike MMMT), and thus is not recommended for assessing alteration in the median annual  
 276 FDC.



277 **Figure 3:** Percentage of time that 117 U. S. dam sites had significant alteration to median annual  
 278 FDC at different FDC percentiles.  
 279

280

281 **3.2 Graphical Illustration of Hydrologic Disturbance at Dam Sites**

282 To better understand the impacts of dam and reservoir disturbances on the hydrologic flow  
283 regime, 4 of the 117 dam sites were selected to further examine hydrologic alteration and to  
284 identify and attribute drivers of the disturbance. Each of the selected sites was paired with a less  
285 regulated site (a nearby site with a lower HDI). The selection criteria for these paired less  
286 regulated sites were: (1) having a concurrent record with the disturbed site, (2) being located in  
287 the same 2-digit US Hydrologic Unit Code (HUC), and (3) having a drainage area within 70% to  
288 130% of the dam site’s drainage area. The purpose of this comparison is to assess whether the  
289 hydrologic disturbance results from anthropogenic activities or due to other drivers (e.g.,  
290 regional climate), in which case an alteration should be evident at both sites. A summary of the  
291 dam and reservoir characteristics at each site can be found in Table 1 (data from USACE, 2000).  
292 The four sites with dams were selected based on different patterns of alteration as shown in the  
293 column labeled “Purpose” in Table 1. Watershed and hydrologic information for both the dam  
294 sites and paired sites can be found in Table 2 (data from Falcone, 2011).

295  
296 One of the primary contributions of this study is to provide a graphical display of the results of  
297 our hypothesis tests which is very easy to understand and communicate. This contribution is  
298 illustrated in Figures 4a-d which document the pre- and post-disturbance annual and median  
299 annual FDCs at both the disturbed and undisturbed sites. The annual FDCs for the pre- and post-  
300 disturbance records are the lighter lines on the plots, while the darker symbols are the median  
301 annual FDCs for the pre- and post-disturbance records. When the p-values of the MMT are  
302 less than 1%, between 1% and 5%, and between 5% and 10%, the points on the median annual  
303 FDCs are colored red, orange and yellow, respectively, and when the p-value is greater than 10%

304 the points are blue. The resulting figures provided a comprehensive graphical tool to assess the  
 305 magnitude of hydrologic alteration across the range of exceedance probabilities.

306 **Table 1:** National Inventory of Dams (NID) information for 4 selected dam/reservoir sites.

Stream	NID ID	H	S	S <sub>Max</sub>	Q <sub>Max</sub>	L	Purpose	Dam Name	Dam Operator
Tualatin River near Dilley, OR	OR03317	22	27.5	27.5	111	300	IP	Tualatin Park	Not provided
Fine Creek at Fine Creek Mills, VA	VA145025	11	19	69	267	415	IR	Telmans Dam	USDA NRCS
Encampment River near Encampment, WY	CO00991	31	38	94	1160	140	FPR	Ginger Quill Reservoir	Ginger Quill Ranch Corp
Priest River near Coolin, ID	ID00318	12	76100	Not provided	4400	194	FR	Priest Lake Outlet Dam	IDWR

307 Variables: **H:** Dam Height (ft), **S:** Normal Storage (acre-ft), **S<sub>Max</sub>:** Maximum Storage (acre-ft),  
 308 **Q<sub>Max</sub>:** Maximum Discharge (cfs), **L:** Dam Length (ft)  
 309 Purpose: **H:** Hydropower, **R:** Recreation, **C:** Flood Control, **I:** Irrigation, **F:** Fish and Wildlife  
 310 Pond, **P:** Fire Protection, Stock, or Small Fish Pond.

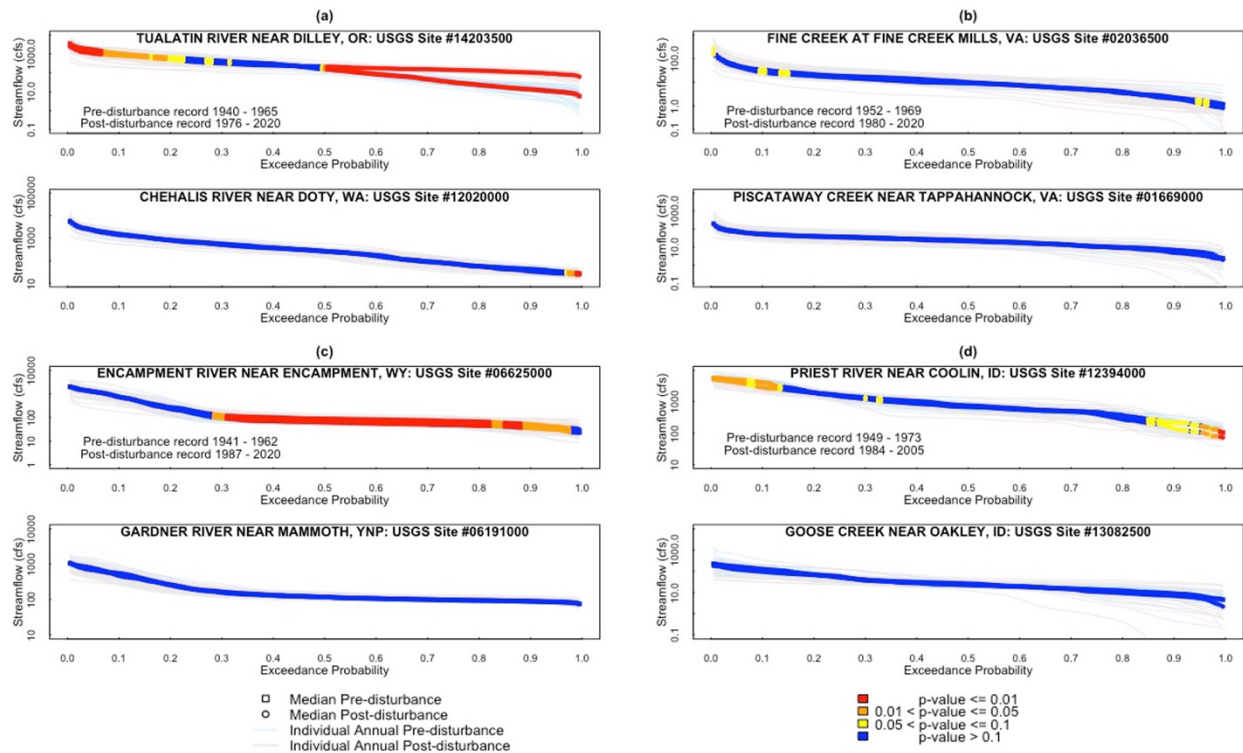
311  
 312  
 313

**Table 2:** Streamflow and watershed information for paired dam and unregulated watersheds.

Stream	USGS ID	Drainage Area (km <sup>2</sup> )	HDI*	Mean Annual Precipitation (cm)	Dam Storage on Stream (MI/km <sup>2</sup> )
Tualatin River near Dilley, OR	14203500	325	25	115	228
Chehalis River near Doty, WA	12020000	293	8	143	0
Fine Creek at Fine Creek Mills, VA	02036500	58	12	110	14.2
Piscataway Creek near Tappahannock, VA	01669000	72.3	6	113	0
Encampment River near Encampment, WY	06625000	678	17	28	41.6
Gardner River near Mammoth, MT	06191000	513	5	35	0
Priest River near Coolin, ID	12394000	1570	11	77	112.29
Goose Creek near Oakley, ID	13082500	1633	7	32.14	0.04

314 \*HDI: Hydrologic Disturbance Index





315

316 **Figure 4:** Graphical Hypothesis Test Results of MMMT at paired dam/no dam sites: a) Tualatin  
 317 River/Chehalis River, b) Fine Creek/Piscataway Creek, c) Encampment River/Gardner River,  
 318 and d) Priest River/Goose Creek. See Tables 1 and 2 for details on selected sites.  
 319

320 Figure 4a shows the pre-and post-disturbance median annual FDCs for the Tualatin River near  
 321 Dilley, OR (USGS#14203500), where a reservoir was placed on the stream in 1970 and the HDI  
 322 is 25. This site was paired with a site subject to minor regulation, the Chehalis River near Doty,  
 323 WA (USGS#12020000) with an HDI of 8. The dam on the Tualatin River provides flood  
 324 protection which significantly lowers high flows and provides drought augmentation which  
 325 significantly increases low streamflows. Alteration in median annual FDC percentiles below an  
 326 exceedance probably of about 5% (extreme high flows) and above an exceedance probability of  
 327 50% (generally low streamflows) were significant at a 1% level. On the Chehalis River,  
 328 significant alterations were only observed for the highest exceedance percentiles (extreme low

329 flows). The Tualatin River has a multipurpose dam operated for flood control, irrigation water  
330 supply, fishing, and fire protection (USACE, 2000).

331  
332 Figure 4b contains the pre- and post-alteration median annual FDCs of Fine Creek at Fine Creek  
333 Mills in Virginia (USGS#02036500), where a reservoir has been in operation since 1974 and has  
334 an HDI of 12. This site was paired with Piscataway Creek near Tappahannock, VA  
335 (USGS#01669000) with an HDI of 6. At Fine Creek only minimal significant alteration ( $5\% < p$ -  
336 value  $< 10\%$ ) was observed at a few percentiles while Piscataway Creek showed no signs of  
337 significant alteration during the same time interval (water years 1944-2020). The minimal  
338 significant alteration at Fine Creek may be due to the relatively small reservoir on this river with  
339 a primary purpose of irrigation water supply and recreational activities (USACE, 2000).

340  
341 Figure 4c shows the pre- and post-alteration median annual FDCs for the Encampment River  
342 near Encampment, WY (USGS#06625000), where a reservoir has been in operation since 1967  
343 and the HDI is 17. This site was paired with a site with minor alteration, the Gardner River near  
344 Mammoth, MT (USGS#06191000) with an HDI of 5. More than two thirds of the median  
345 annual FDC of the Encampment River (intermediate and low flows) had a significant alteration  
346 at a 1% level. In contrast, the FDC of the Gardner River indicates no significant alteration  
347 through the same time period (water years 1940-2020). This could suggest that the alteration in  
348 the FDC of the Encampment River was driven by a reservoir created primarily for a fish and  
349 wildlife pond, fire protection and recreation (USACE, 2000). Unlike the Tualatin River (Figure  
350 4a), no significant alteration to the high flows on the Encampment River was observed. Unlike  
351 the Encampment River, the Tualatin River reservoir had an additional operating goal to provide  
352 downstream flood protection.

353  
354 Figure 4d illustrates the median annual FDCs for the Priest River near Coolin, ID  
355 (USGS#12394000), where a reservoir was placed on the stream in 1978 with an HDI of 11. This  
356 site was paired with a site with minor regulation, Goose Creek near Oakley, ID  
357 (USGS#13082500) with an HDI of 7. Both low flows and extreme high flows have been  
358 significantly altered on the Priest River. Median annual FDC percentiles below an exceedance  
359 probably of about 5% (extreme high flows) and median annual FDC percentiles above an  
360 exceedance probability of 95% were significant at a 5% level. On Goose Creek, there was no  
361 significant alteration to the median annual FDC during the same time period (water years 1949-  
362 2005). The significant alteration to low flows on the Priest River was a decrease and not an  
363 increase in low flows, which is commonly expected when a dam is used to augment streamflow  
364 during drought periods. Priest Lake maintains high water levels during the summer months for  
365 recreational purposes, and then drops the water level in the fall to accommodate the spawning of  
366 Kokanee salmon (BCLC, 2021). Such regulation appears to have caused a drop in low  
367 streamflows after the dam was installed. In 2018 the Idaho Water Resources Board started the  
368 “Priest Lake Water Management Project” to upgrade the Priest Lake outlet dam to improve low  
369 flow management (IDWR, 2021).

### 370 **3.3. Graphical Illustration of Other Forms of Hydrologic Disturbance**

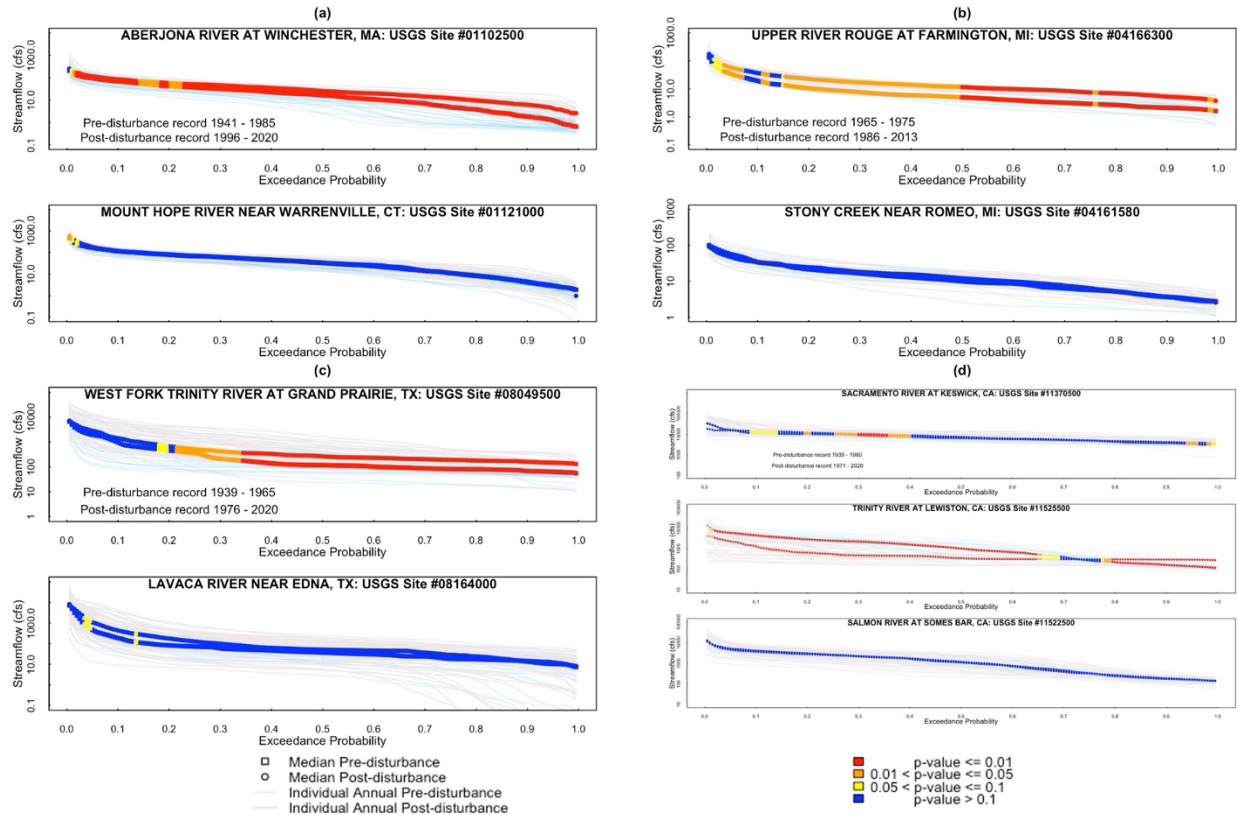
371 To examine the impacts of different forms of hydrologic disturbance on median annual FDCs,  
372 five sites were identified: two sites that have undergone urbanization activities, one site  
373 potentially disturbed by wastewater treatment plant (WWTP) effluent discharge, and two sites  
374 that have undergone an interbasin water transfer (source and receiving end). The first site  
375 examined was the Aberjona River at Winchester, MA (USGS#01102500) with an HDI of 17.

376 This site was paired with a site with minor regulation, the Mount Hope River near Warrentville,  
377 CT (USGS#01121000) with an HDI of 12. The Aberjona River has undergone an increasing  
378 trend in urbanization over time; for this analysis we chose an approximate midpoint of the  
379 historical record as the year of alteration (1990). As shown in Figure 5a, the entire FDC of the  
380 Aberjona River witnessed an alteration with more than 90% of the quantiles showing significant  
381 alteration at a 1% significance level and the remaining percentiles showing significant alteration  
382 at a 5% significance level. The Mount Hope River only exhibited alteration for the smallest FDC  
383 exceedance percentiles (largest flows) over the concurrent record (water years from 1940-2020).  
384 The alteration on the Aberjona River is clearly linked to urbanization activities, with the area  
385 experiencing a gradual change in land use, from primarily farms to residential and urban areas  
386 (Solo-Gabriele and Perkins, 1997). This result is also consistent with the findings of Allaire et  
387 al. (2015) who provide additional information on the impact of urbanization and water  
388 withdrawals on FDCs of the Aberjona River.

389

390 The second urbanization site is the Upper River Rouge at Farmington, MI (USGS#04166300)  
391 with an HDI of 15 (Figure 5b). This site was paired with a site with a slightly lower HDI, Stony  
392 Creek at Romeo, MI (USGS#04161580) with an HDI of 12. Nearly all of the FDC percentiles of  
393 the Upper River Rouge shows a significant alteration with flows with greater than a 50%  
394 exceedance probability (low flows) significant at a 1% level. The absence of alteration at the  
395 reference site led us to link alteration on the Upper River Rouge to the documented increase in  
396 housing density in the watershed (Falcone, 2017).

397



398  
399

400 **Figure 5:** Results of MMMT at paired disturbed/not disturbed sites: a) Aberjona River  
 401 (urbanization)/Mount Hope River, Upper River Rouge (urbanization)/Stony Creek, (b) West  
 402 Fork (WF) Trinity River (wastewater effluent)/Lavaca River, and (d) Sacramento, Trinity River  
 403 (interbasin transfer)/Salman River.

404

405 Figure 5c contains the pre- and post-alteration FDCs on the West 1 Trinity River at  
 406 Grand Prairie, TX (USGS#08049500) with an HDI of 23. Upstream from this site, the Village  
 407 Creek Water Reclamation Facility discharges wastewater effluent into the stream. This site was  
 408 paired with the Lavaca River near Edna, TX (USGS#08164000) with an HDI of 13. On the West  
 409 Fork Trinity River, almost two thirds of the FDC percentiles (intermediate and low flows) show  
 410 significant alteration at a 1% level, while at the Lavaca River for the same concurrent record  
 411 (water years from 1939-2020) a significant alteration at a 10% significance level occurred at only  
 412 a few (high flow) percentiles near an exceedance probability of 1%. This example documents the  
 413 significant impact wastewater effluent can have on streamflow, especially for moderate to low

414 flows. One would expect such impacts to be related to the magnitude of both the wastewater  
415 discharge and streamflow.

416

417 Figure 5d contains pre- and post-alteration annual FDCs for three rivers. The first two are the  
418 Sacramento River at Keswick, CA (USGS#11370500; HDI 16) and the Trinity River at  
419 Lewiston, CA (USGS#11525500; HDI 17), with the Sacramento River receiving water  
420 transferred from the Trinity River. These two sites were paired with a site with very minor  
421 regulation, the Salmon River at Somes Bar, CA (USGS#11522500) with an HDI of 6. The  
422 Central Valley Project utilizes the surplus water (704,000 acre-feet/year) in the Trinity River  
423 Basin for the benefit of the San Joaquin River Valley. While this water transfer only has minimal  
424 significant impact on the Sacramento River with some significant changes between the  
425 exceedance probabilities of 10% and 40% and above 95%, on the Trinity River a much more  
426 significant impact can be observed, with median annual FDC percentiles significant at a 1% level  
427 across nearly the entire FDC. In contrast, the FDC of the minimally regulated Salmon River,  
428 showed no signs of significant alteration. Interbasin transfers are not the only form of alteration  
429 on the Sacramento and Trinity Rivers, as both sites have upstream lakes/reservoirs (Lake Shasta  
430 and Trinity Lake, respectively). The hydrologic alteration exhibited by these two rivers could be  
431 related to the combined impact of the interbasin water transfer as well as upstream reservoir  
432 operations.

### 433 **3.4 HCDN Sites**

434 McCabe and Wolock (2002) found a significant increase in average annual minimum and  
435 median daily streamflow at a subset of HCDN sites in the CONUS starting in 1970, with the  
436 most prominent increases in the eastern U.S. They attributed these increases to step changes in

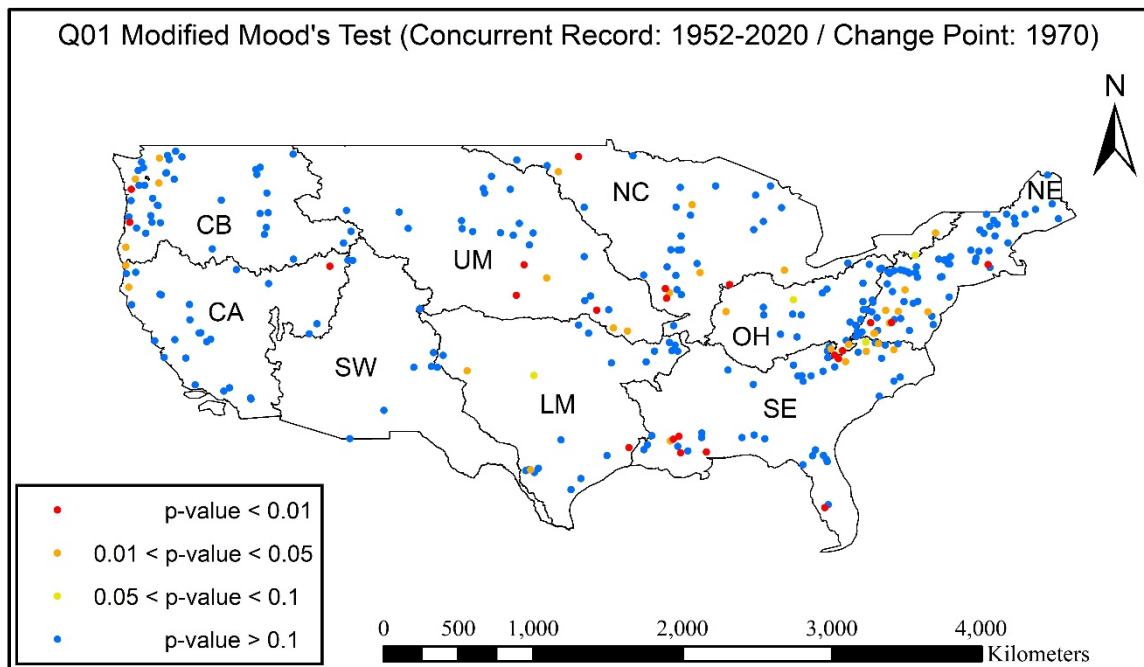
437 precipitation around 1970, a result corroborated by Small et al. (2006), Karkauer and Fung  
438 (2008) and Wang and Hejazi (2012). Of interest here is applying the MMMT to minimally  
439 regulated streamflow sites in the CONUS using the 1970 breakpoint to assess potential climatic  
440 changes on streamflows across the U.S. Here we employ a subset of the USGS's HCDN-2009  
441 watersheds, which are generally considered minimally regulated, to assess significant changes in  
442 the median annual FDC at exceedance probabilities of 1% (Q01, high flows), 50% (Q50, median  
443 flows), and 99% (Q99, low flows) using the 1970 break point. Only HCDN-2009 sites with at  
444 least 15 years before 1970 and 15 years after 1970 were employed. To maximize the concurrent  
445 record our pre-disturbance record was 1952 – 1970 and our post-disturbance record was 1971 –  
446 2020, resulting in 308 HCDN sites. We employ concurrent records across all sites so that our  
447 results are less impacted by varying record lengths and the regional climate is also expected to be  
448 more consistent across sites.

449

450 Figure 6 illustrates the results for Q01 (high flows), where the HCDN site locations are colored  
451 red if the p-value is less than 0.01, yellow if between 0.01 and 0.05, and orange if between 0.05  
452 and 0.1 and blue otherwise. Since most of the points in Figure 5 are blue, we conclude that there  
453 are only minimum significant changes in the median annual Q01 before and after 1970. Table 3  
454 contains the number of sites in each super region with significant changes in the median annual  
455 Q01. The percentage of sites with significant changes in the median annual Q01 ( $p$ -value  $< 0.1$ )  
456 ranged from 0% to 31% across super regions, with the upper midwestern U.S. and the  
457 southeastern U.S. having the largest percentages of significant differences.

458

459 Figure 7 illustrates the results for Q99 (low flows). Unlike for Q01, we see many significant  
 460 changes in the median annual Q99. The percentage of sites with significant changes in the  
 461 median annual Q99 (p-value < 0.1) ranged from 18% to 81% across super regions, with the  
 462 upper midwestern U.S. having the most sites with significant changes and the midwestern and  
 463 eastern U.S. having more than 40% of the sites with significant changes in the average Q99.  
 464 These results are similar to those found by Douglas et al. (2000) who examined regional  
 465 significance of trends in high and low flows, finding the upper midwestern U.S. to have the most  
 466 significant trends in low flows. They are also similar to results found by McCabe and Wolock  
 467 (2002) who identified significant changes in low streamflows in the northeastern U.S. using a  
 468 1970 break point.  
 469



470 **Figure 6:** Results of MMT for high flows (Q01) at 308 HCDN sites.  
 471

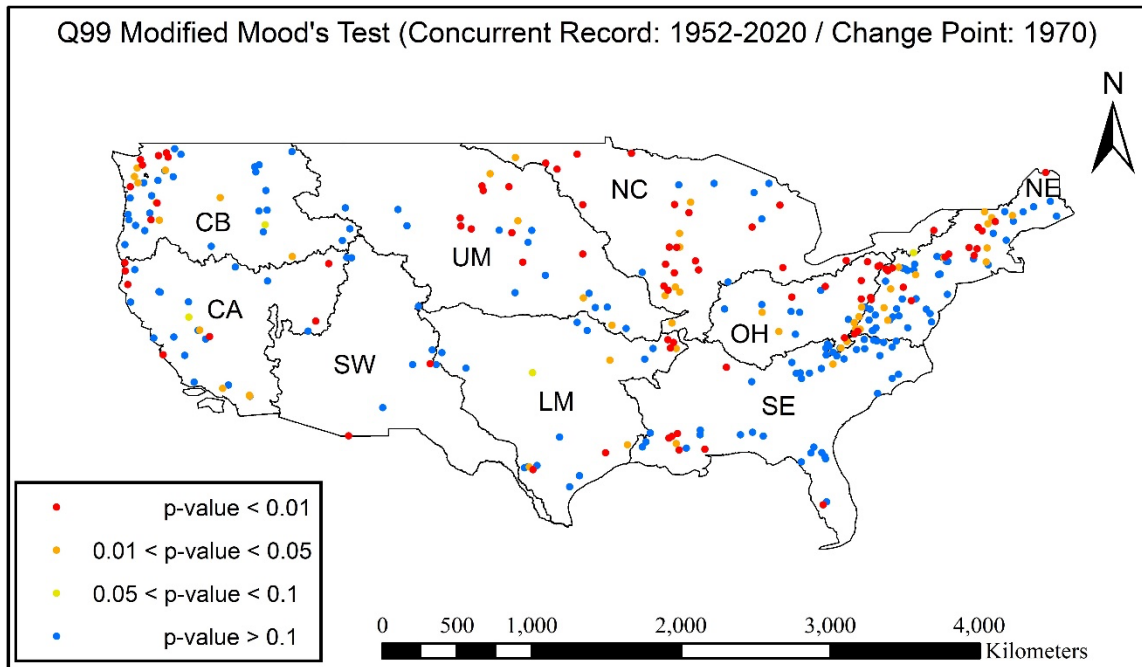
472



473 **Table 3:** Percentage of sites with significant changes in the median annual Q01, Q99 and Q50.

Region	Number of Sites	High Flows (Q01)				Median Flows (Q50)				Low Flows (Q99)			
		p < 0.01	0.01 < p < 0.05	0.05 < p < 0.1	p > 0.1	p < 0.01	0.01 < p < 0.05	0.05 < p < 0.1	p > 0.1	p < 0.01	0.01 < p < 0.05	0.05 < p < 0.1	p > 0.1
NE	70	4.3%	10%	0.0%	85.7%	40.0%	30.0%	0.0%	30.0%	24.3%	20.0%	0.0%	55.7%
OH	26	7.7%	7.7%	7.7%	76.9%	61.5%	23.1%	3.8%	11.5%	34.6%	26.9%	0.0%	38.5%
SE	50	12.0%	12.0%	2.0%	74.0%	20.0%	18.0%	2.0%	60.0%	14.0%	4.0%	0.0%	82.0%
NC	36	11.1%	16.7%	2.8%	69.4%	41.7%	30.6%	2.8%	25.0%	55.6%	22.2%	2.8%	19.4%
UM	25	12.0%	12.0%	0.0%	76.0%	32.0%	12.0%	0.0%	56.0%	36.0%	16.0%	0.0%	48.0%
LM	23	4.3%	8.8%	4.3%	82.6%	17.4%	21.8%	4.3%	56.5%	21.8%	17.4%	4.3%	56.5%
SW	8	0.0%	0.0%	0.0%	100%	0.0%	25.0%	0.0%	75.0%	25.0%	0.0%	0.0%	75.0%
CB	42	4.8%	9.5%	0.0%	85.7%	2.4%	4.8%	0.0%	92.9%	19.0%	19.0%	2.4%	59.5%
CA	28	3.6%	7.1%	0.0%	89.3%	10.7%	3.6%	0.0%	85.7%	25.0%	14.3%	3.6%	57.1%
<b>Total</b>	<b>308</b>												

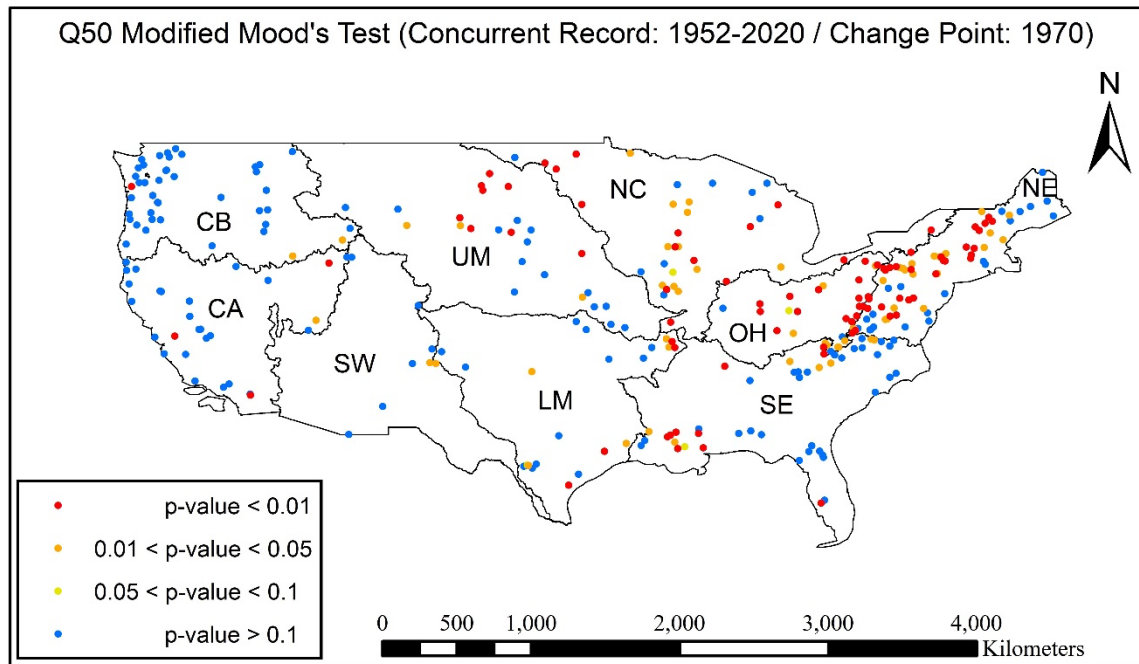
474



475 **Figure 7:** Results of MMMT for low flows (Q99) at 308 HCDN sites.  
476

477  
478 Figure 8 illustrates results for median annual Q50 (median flows), which indicate significant  
479 changes in the median annual Q50 at many HCDN sites across the U.S. The percentage of sites  
480 with significant changes in the median annual Q50 (p-value < 0.1) ranged from 7% to 88%  
481 across super regions, with the upper midwestern U.S., Ohio River Region and eastern U.S.

482 having the highest percentage of sites with significant changes in the median annual Q50. These  
483 results are similar to those found by McCabe and Wolock (2002), who also identified significant  
484 changes in the eastern U.S.



485  
486 **Figure 8:** Results of MMT for median flows (Q50) at 308 HCDN sites.

487  
488 **4.0 Discussion**

489 Our graphical FDC-based hypothesis test based on MMT has enabled us to demonstrate and  
490 illustrate that the alteration of flow regimes varies with the form and magnitude of disturbances.  
491 One would expect that alterations caused by a dam and reservoir operations would be different  
492 from alterations resulting from other forms of disturbance such as urbanization, climate change  
493 and/or wastewater treatment plant effluent discharge. Even for sites with a dam and reservoir,  
494 there was large variability in the alteration patterns. At a 5% significance level, around 41% of  
495 the sites with dams had significant alteration only at low flows, 10% of the dam sites showed  
496 significant alteration for only the high flows, and approximately 28% of the dam sites saw  
497 significant alterations for both low and high flows. While examining the reservoir operations

498 policy is not part of this study, one might expect that the level of hydrologic alteration could be  
499 related to operating rules, characteristics of the reservoir and the primary purpose(s) of the  
500 reservoir system. Of course, there is also the possibility of several different disturbances acting  
501 together at a site. While the Tualatin River downstream from a multipurpose dam showed the  
502 expected behavior of significant alteration to high flows (decreasing) and low flows (increasing),  
503 the other three selected dam sites showed varied patterns of alteration. These patterns appear to  
504 be linked to the characteristics of the reservoirs and their management policies and purposes,  
505 though other additional forms of disturbance may also be involved at these sites.

506

507 Sites with other forms of hydrologic disturbance were also examined. Urbanization caused  
508 significant alterations (increases) to the entire FDC, a result consistent with the study by Allaire  
509 et al. (2015). At the West Fork Trinity River, where discharge from a wastewater treatment plant  
510 occurs above the gauge, two thirds of the FDC percentiles (intermediate and low flows) showed  
511 a significant alteration. This significant alteration for low and intermediate flows was not a  
512 surprise, given that the river may be composed of up to 90% wastewater during dry months  
513 (VCWRF, 2021). The impact of interbasin water transfers depends on the magnitude of both the  
514 transfer and the streamflows at both the donor and receiving watersheds. Additional sites with  
515 various forms of potential hydrologic disturbance should be examined to draw broader  
516 conclusions regarding patterns of alteration and other tools useful for evaluating the impact of  
517 multiple disturbances on streamflow regimes could also prove useful such as the multivariate  
518 elasticity approach introduced by Allaire et al. (2015) as well as the water regimes approach  
519 introduced by Weiskel et al. (2007).

520

521 Application of our methodology at minimally regulated HCDN sites spatially distributed across  
522 the CONUS indicated significant alteration to median and low streamflows in many regions of  
523 the U.S. when a 1970 break point was employed. The upper midwestern U.S. and eastern U.S.  
524 had the greatest densities of HCDN sites with significant alteration to median annual FDC  
525 percentiles with exceedance probabilities of 50% (median flows) and 99% (low flows). These  
526 results are consistent with those found by McCabe and Wolock (2002), who identified step  
527 changes in precipitation patterns in the eastern U.S. which drove changes in low flows before  
528 and after 1970. Douglas et al. (2000) also identified regionally significant trends in low flows in  
529 the upper midwestern U.S., which is consistent with our results. We recommend the application  
530 of the MMT considering other break points, to further explore and illustrate the impacts of  
531 drivers of changes in hydrologic patterns across the U.S.

532

## 533 **5.0 Conclusions and Recommendations**

534 There is a need for a convenient, simple to understand graphical tool that illustrates and evaluates  
535 the statistical significance of impacts of hydrologic alteration across the entire range of  
536 streamflow. This study explored the use of Modified Mood's Median Test (MMMT), a non-  
537 parametric hypothesis tests that relies on no distributional or shape assumptions, for evaluating  
538 the significance and degree of hydrologic alteration based on the median annual flow duration  
539 curve (FDC). Importantly, this rigorous hypothesis test developed by others (Brown and Mood,  
540 1951; Fligner and Rust, 1982) is introduced, evaluated, and converted into a graphical hydrologic  
541 tool to easily assess and illustrate hydrologic alteration across a wide variety of daily streamflow  
542 conditions. Prior to a potential hydrologic disturbance, it is assumed that there is an existing  
543 unaltered period of record with which the post-disturbance altered streamflow record can be

544 compared. The graphical hypothesis test introduced was evaluated at 117 gaged sites spatially  
545 distributed across the U.S. (Figure 1) where a dam and reservoir has been placed on the stream,  
546 and 5 additional sites with other forms of hydrologic disturbance (urbanization, wastewater  
547 treatment plant discharge, and interbasin water transfer). Different types of disturbance generally  
548 alter the streamflow regime in different ways. In addition, using minimally regulated HCDN  
549 sites, we also evaluate this test to assess the hydrologic impacts of regional climate variations.

550

551 The results of these experiments indicated that:

- 552 • At sites with dams, significant hydrologic alteration was generally found to be much  
553 more likely for low flows than for flood flows. Across 117 sites with dams, Figure 3  
554 documents that the percentage of sites with significant alteration consistently increased as  
555 the exceedance probability increased (flow magnitudes decreased). In contrast, at the  
556 lowest exceedance probabilities (flood flows) the percentage of sites with significant  
557 alteration consistently decreased as their flow exceedance decreased (flow increased).
- 558 • The four dam sites selected for more detailed analysis showed different patterns of  
559 alteration over different portions of the FDCs. These different patterns were generally  
560 linked to the location, characteristics (e.g., storage volume) and the primary purpose of  
561 the dam and reservoir.
- 562 • The two sites disturbed by urbanization showed significant alterations at nearly all FDC  
563 percentiles, indicating urbanization activities generally increase all magnitudes of  
564 streamflow.
- 565 • Sites disturbed by effluent discharge from a wastewater treatment plant generally resulted  
566 in significant alteration at intermediate and low flows.

- 567 • At sites with interbasin water transfers, the significance of the alteration appears to be a  
568 function of the magnitude of both the streamflows at the sites and the characteristics of  
569 the water transfers.
- 570 • At HCDN sites, when using the year 1970 as a breakpoint, there are significant changes  
571 to low and median flows, particularly in midwestern and northeastern U.S., yet our  
572 results indicate only minimal significant change in high flows.

573

574 We have developed a convenient, general, and simple to understand graphical tool to assess  
575 hydrologic alteration. The tool requires the presence of a pre- and post-alteration hydrologic  
576 series and assesses whether hydrologic flows in a “typical” (median) year have significantly  
577 changed across a wide variety of hydrologic conditions. The MMT is a non-parametric test  
578 that requires no assumption regarding the distribution of the pre- and post-alteration series.

579 Importantly, the graphical hypothesis testing tool introduced here is very general, and although  
580 we have only applied it to daily streamflow series in this analysis, it could be employed with any  
581 temporal series to assess system changes. Furthermore, the test is suited to any application of  
582 FDC’s, one of the most widely used graphical instruments employed in the field of hydrology.

583 For example, FDC’s can accommodate seasonal variations by developing FDC’s by month or  
584 season. Similarly, Vogel and Fennessey (1995) document how FDC’s can be applied to any  
585 “water resource index derived from streamflow such as hydropower plant power output, instream  
586 pollutant concentrations and/or loads, and ecological habitat suitability indices.” It is our hope  
587 that this graphical hypothesis test can serve as a graphical user interface in a wide range of water  
588 resource software applications where FDC’s provide an overall summary metric of watershed

589 system behavior. This tool, MMMT, is available on Github at  
590 <https://github.com/cnkroll/MMMT.git>.

## 591 **Acknowledgements**

592 The authors would like to acknowledge the support received from the Amideast Fulbright  
593 Foreign Student Program and the SUNY College of Environmental Science and Forestry in  
594 Syracuse, NY.

## 595 596 **6.0 References**

- 597  
598 Allaire, M.C., Vogel, R.M., Kroll, C.N., 2015. The hydromorphology of an urbanizing watershed  
599 using multivariate elasticity. *Advances in Water Resources* 86, 147-154.  
600  
601 Allan, J.D., 2004. Landscapes and riverscapes: the influence of land use on stream ecosystems.  
602 *Annu Rev Ecol Evol Syst* 35, 257–284.  
603  
604 Ames, D.P., 2006. Estimating 7Q10 confidence limits from data: a bootstrap approach. *J. Water*  
605 *Resour. Plan. Manag.* 132, 204–208.  
606  
607 Atieh, M., Taylor, G., Sattar, A.M., Gharabaghi, B., 2017. Prediction of flow duration curves for  
608 ungauged basins. *J. Hydrol.* 545, 383–394.  
609  
610 Blum, A.G., Archfield, S.A., Vogel R.M., 2017. The probability distribution of daily streamflow  
611 in the United States. *Hydrol. Earth Syst. Sci.* 21, 3093–3103.  
612  
613 Bonett, D.G., Price Jr, R.M., 2020. Confidence Intervals for Ratios of Means and Medians. *J.*  
614 *Educ. Behav. Stat.* 45, 750–770.  
615  
616 Bonner County Lakes Commission [Internet]. Priest River (ID): [cited 2021 APR 25]. Available  
617 from: <https://lakescommission.wordpress.com/priest-lake>.  
618  
619 Bradford, J.B., Betancourt, J.L., Butterfield, B.J., Munson, S.M., Wood, T.E., 2018. Anticipatory  
620 natural resource science and management for a changing future. *Front. Ecol. Environ.* 16,  
621 295–303.  
622  
623 Brown, A.E., Lu, Z., McMahon, T.A., Western, A.W., Vertessy, R.A., 2005. A review of paired  
624 catchment studies for determining changes in water yield resulting from alterations in  
625 vegetation. *J Hydrol* 310(1–4), 28–61.  
626

627 Brown, G.W. and Mood, A.M., 1951. On median tests for linear hypotheses. In Proceedings of  
628 the Second Berkely Symposium on Mathematical Statistics and Probability (pp. 159-  
629 166). University of California Press.  
630

631 Bunn, S.E., Arthington, A.H., 2002. Basic principles and ecological consequences of altered  
632 flow regimes for aquatic biodiversity. *Environ. Manage.* 30, 492–507.  
633

634 Castellarin, A., Botter, G., Hughes, D.A., Liu, S., Ouarda, T., Parajka, J., Post, D.A., Sivapalan,  
635 M., Spence, C., Viglione, A., 2013. Prediction of flow duration curves in ungauged  
636 basins. *Runoff Prediction in Ungauged Basins Synth. Process. Places Scales*, 135–162.  
637

638 Castellarin, A., Camorani, G., Brath, A., 2007. Predicting annual and long-term flow-  
639 duration curves in ungauged basins. *Advances in Water Resources* 30(4), 937-953.  
640

641 Dudley, R.W., Hirsch, R.M., Archfield, S.A., Blum, A.G., Renard, B., 2020. Low streamflow  
642 trends at human-impacted and reference basins in the United States. *J. Hydrol.* 580,  
643 124254.  
644

645 Ernst, A.F., Albers, C.J., 2017. Regression assumptions in clinical psychology research  
646 practice—a systematic review of common misconceptions. *PeerJ* 5, e3323.  
647

648 Falcone, J.A., 2011. GAGES-II: Geospatial attributes of gages for evaluating streamflow. US  
649 Geological Survey.  
650

651 Falcone, J.A., 2017. US Geological Survey GAGES-II time series data from consistent sources  
652 of land use, water use, agriculture, timber activities, dam removals, and other historical  
653 anthropogenic influences: US Geological Survey data release. In US Geological Survey  
654 Data Release.  
655

656 Falcone, J.A., Carlisle, D.M., Wolock, D.M. Meador, M.R., 2010. GAGES: A stream gage  
657 database for evaluating natural and altered flow conditions in the conterminous United  
658 States: Ecological archives E091-045. *Ecology* 91(2), 621-621.  
659

660 Fligner, M.A., Rust, S.W., 1982. A modification of Mood’s median test for the generalized  
661 Behrens-Fisher problem. *Biometrika* 69, 221–226.  
662

663 Friedman, M., 1937. The use of ranks to avoid the assumption of normality implicit in the  
664 analysis of variance. *J. Am. Stat. Assoc.* 32, 675–701.  
665

666 Gao, Y., Vogel, R.M., Kroll, C.N., Poff, N.L., Olden, J.D., 2009. Development of representative  
667 indicators of hydrologic alteration. *J. Hydrol.* 374, 136–147.  
668

669 Hecht, J.S., Cai, X., Eheart, J.W., 2014. Operating rules for an off-stream blending reservoir to  
670 control nitrate in a municipal water system. *J. Water Resour. Plan. Manag.* 140,  
671 04014015.  
672



673 Hecht, J.S., Vogel, R.M., McManamay, R.A., Kroll, C.N., Reed, J.M., 2020. Decision Trees for  
674 Incorporating Hypothesis Tests of Hydrologic Alteration into Hydropower–Ecosystem  
675 Tradeoffs. *J. Water Resour. Plan. Manag.* 146, 04020017.  
676

677 Helsel, D.R., Hirsch, R.M., Ryberg, K.R., Archfield, S.A., Gilroy, E.J., 2020, Statistical methods  
678 in water resources: U.S. Geological Survey Techniques and Methods, book 4, chap. A3,  
679 458 p.  
680

681 Idaho Department of Water Resources [Internet]. Priest Lake Water Management Project (ID):  
682 [cited 2021 APR 25]. Available from: <https://idwr.idaho.gov/IWRB/projects/priest-lake/>  
683

684 Jumani, S., Deitch, M.J., Kaplan, K., Anderson, E.P., Krishnaswamy, J., Lecours, V., Whiles,  
685 M.R., 2020. River fragmentation and flow alteration metrics: a review of methods and  
686 directions for future research, *Env. Res. Let.*, 15(12), 123009.  
687

688 Kroll, C.N., Croteau, K.E., Vogel, R.M., 2015. Hypothesis tests for hydrologic alteration. *J.*  
689 *Hydrol.* 530, 117–126.  
690

691 Leong, C. and Yokoo, Y., 2021. A step toward global-scale applicability and transferability of  
692 flow duration curve studies: A flow duration curve review (2000–2020). *Journal of*  
693 *Hydrology*, 603, p.126984.  
694

695 Magilligan, F.J., Nislow, K.H., 2005. Changes in hydrologic regime by dams. *Geomorphology*  
696 71, 61–78.  
697

698 Malmqvist, B., Rundle, S., 2002. Threats to the running water ecosystems of the world. *Environ.*  
699 *Conserv.* 134–153.  
700

701 Moyle, P.B., Mount, J.F., 2007. Homogenous rivers, homogenous faunas. *Proc. Natl. Acad. Sci.*  
702 104, 5711–5712.  
703

704 Nilsson, C., Reidy, C.A., Dynesius, M., Revenga, C., 2005. Fragmentation and flow regulation  
705 of the world’s large river systems. *Science* 308, 405–408.  
706

707 Olden, J.D., Poff, N.L., 2003. Redundancy and the choice of hydrologic indices for  
708 characterizing streamflow regimes. *River Res. Appl.* 19, 101–121.  
709

710 Poff, N.L., Olden, J.D., Merritt, D.M., Pepin, D.M., 2007. Homogenization of regional river  
711 dynamics by dams and global biodiversity implications. *Proc. Natl. Acad. Sci.* 104,  
712 5732–5737.  
713

714 Poff, N.L., Richter, B.D., Arthington, A.H., Bunn, S.E., Naiman, R.J., Kendy, E., Acreman, M.,  
715 Apse, C., Bledsoe, B.P., Freeman, M.C., 2010. The ecological limits of hydrologic  
716 alteration (ELOHA): a new framework for developing regional environmental flow  
717 standards. *Freshw. Biol.* 55, 147–170.  
718

719 Postel, S.L., Daily, G.C., Ehrlich, P.R., 1996. Human appropriation of renewable fresh water.  
720 Science 271, 785–788.  
721

722 Richter, B.D., Baumgartner, J.V., Powell, J., Braun, D.P., 1996. A method for assessing  
723 hydrologic alteration within ecosystems. *Conserv. Biol.* 10, 1163–1174.  
724

725 Rosenberg, D.M., McCully, P., Pringle, C.M., 2000. Global-scale environmental effects of  
726 hydrological alterations: introduction. *BioScience* 50, 746–751.  
727

728 Sabater, S., Bregoli, F., Acuña, V., Barceló, D., Elosegi, A., Ginebreda, A., Marcé, R., Muñoz,  
729 I., Sabater-Liesa, L., Ferreira, V., 2018. Effects of human-driven water stress on river  
730 ecosystems: a meta-analysis. *Sci. Rep.* 8, 1–11.  
731

732 Serinaldi, F., 2011. Analytical confidence intervals for index flow flow duration curves. *Water*  
733 *Resources Research*, 47(2), W02542.  
734

735 Solo-Gabriele, H.M., Perkins, F.E., 1997. Streamflow and suspended sediment transport in an  
736 urban environment. *J. Hydraul. Eng.* 123, 807–811.  
737

738 USACE, N., 2000. U.S. Army Corps of Engineers National Inventory of Dams.  
739

740 Village Creek Water Reclamation Facility [Internet]. Fort Worth (TX): [cited 2021 APR 1].  
741 Available from: <https://www.fortworthtexas.gov/departments/water/village-creek>  
742

743 Vogel, R.M., Fennessey, N.M., 1994. Flow-duration curves. I: New interpretation and  
744 confidence intervals. *J. Water Resour. Plan. Manag.* 120, 485–504.  
745

746 Vogel, R.M., Fennessey, N.M., 1995. Flow duration curves II: A review of applications in water  
747 resources planning I. *JAWRA J. Am. Water Resour. Assoc.* 31, 1029–1039.  
748

749 Vogel, R.M., Sieber, J., Archfield, S.A., Smith, M.P., Apse, C.D., Huber-Lee, A., 2007.  
750 Relations among storage, yield, and instream flow. *Water Resour. Res.* 43, W05403.  
751

752 Vogel, R.M., Lall, U., Cai, X., Rajagopalan, B., Weiskel, P.K., Hooper, R.P. and Matalas, N.C.,  
753 2015. Hydrology: The interdisciplinary science of water. *Water Resources*  
754 *Research*, 51(6), 4409-4430.  
755

756 Wainer, H., 1990. Graphical Visions from William Playfair to John Tukey, *Statistical Science*,  
757 5(3), 340-346.  
758

759 Weiskel, P.K., Vogel, R.M., Steeves, P.A., Zarriello, P.J., DeSimone, L.A., Ries III, K.G., 2007.  
760 Water-use regimes: Characterizing direct human interaction with hydrologic systems.  
761 *Water Resources Research*, 43, W04402.  
762  
763

764 Williams, M.N., Grajales, C.A.G., Kurkiewicz, D., 2013. Assumptions of multiple regression:  
765 Correcting two misconceptions. *Pract. Assess. Res. Eval.* 18, 11.  
766

767 Yang, T., Cui, T., Xu, C.-Y., Ciais, P., Shi, P., 2017. Development of a new IHA method for  
768 impact assessment of climate change on flow regime. *Glob. Planet. Change* 156, 68–79.  
769

770 Yang, Y.-C.E., Cai, X., Herricks, E.E., 2008. Identification of hydrologic indicators related to  
771 fish diversity and abundance: A data mining approach for fish community analysis.  
772 *Water Resour. Res.* 44, W04412.  
773

774 Ye, L., Gu, X., Wang, D. and Vogel, R.M., 2021. An unbiased estimator of coefficient of  
775 variation of streamflow. *Journal of Hydrology*, 594, p.125954.  
776

777 Yusa, A., Berry, P., J Cheng, J., Ogden, N., Bonsal, B., Stewart, R., Waldick, R., 2015. Climate  
778 change, drought and human health in Canada. *Int. J. Environ. Res. Public. Health* 12,  
779 8359–8412.  
780